

Cable bacteria generate a firewall against euxinia in seasonally hypoxic basins

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Seasonal oxygen depletion (hypoxia) in coastal bottom waters can lead to the release and persistence of free sulfide (euxinia), which is highly detrimental to marine life. Although coastal hypoxia is relatively common, reports of euxinia are less frequent, which suggests that certain environmental controls can delay the onset of euxinia. However, these controls and their prevalence are poorly understood. Here we present field observations from a seasonally hypoxic marine basin (Grevelingen, The Netherlands), which suggest that the activity of cable bacteria, a recently discovered group of sulfur-oxidizing microorganisms inducing long-distance electron transport, can delay the onset of euxinia in coastal waters. Our results reveal a remarkable seasonal succession of sulfur cycling pathways, which was observed over multiple years. Cable bacteria dominate the sediment geochemistry in winter, whereas, after the summer hypoxia, *Beggiatoaceae* mats colonize the sediment. The specific electrogenic metabolism of cable bacteria generates a large buffer of sedimentary iron oxides before the onset of summer hypoxia, which captures free sulfide in the surface sediment, thus likely preventing the development of bottom water euxinia. As cable bacteria are present in many seasonally hypoxic systems, this euxinia-preventing firewall mechanism could be widely active, and may explain why euxinia is relatively infrequently observed in the coastal ocean.

sediment biogeochemistry | cable bacteria | coastal hypoxia | sulfur cycling | microbial competition

The depletion of oxygen in bottom waters (hypoxia) is a naturally recurring phenomenon in some coastal systems (1), such as basins with restricted water circulation (2), and shelf regions subject to strong nutrient upwelling (3). Alongside this natural hypoxia, there is evidence for a global increase in the frequency, extent, intensity, and duration of coastal hypoxia, which is linked to an increased anthropogenic input of nutrients into the coastal ocean in combination with climate change (1, 4–6). The development of bottom water hypoxia has major consequences for the functioning of coastal ecosystems, sometimes leading to the formation of “dead zones” characterized by a complete absence of benthic fauna and fish. Areas sensitive to hypoxia are typically major fishing grounds, so the resulting economic and biodiversity losses make the global expansion of coastal hypoxia a subject of growing concern (7, 8).

The ecosystem impacts of coastal hypoxia are particularly amplified when bottom water oxygen depletion progresses to a critical transition, termed euxinia, when free sulfide escapes from the sediment and accumulates in the bottom water (9). Even low levels of free sulfide are toxic to metazoan life, and therefore euxinia can induce mass mortality events, even among highly motile fauna like fish and large crustaceans (9–11). Although strong oxygen depletion, or even a complete removal of oxygen (anoxia), is often reported, concomitant reports of euxinia in coastal bottom waters are much scarcer. This suggests that certain sedimentary processes delay the onset of euxinia relative to anoxia, but, at

present, the environmental controls on the timing and formation of coastal euxinia are poorly understood.

Here we document a microbial mechanism that can delay or even prevent the development of euxinia in seasonally hypoxic basins. The mechanism is based on the metabolic activity of a newly discovered type of electrogenic microorganism, named cable bacteria (*Desulfobulbaceae*, Deltaproteobacteria), which are capable of inducing electrical currents over centimeter-scale distances in the sediment (12, 13). Cable bacteria have recently been suggested to be abundant in seasonally hypoxic coastal systems (14), but their impact on the biogeochemical cycling in these systems is unknown. These filamentous bacteria possess a unique respiratory metabolism in which the oxidative removal of sulfide in deeper sediment layers is electrically coupled to the reductive consumption of oxygen just below the sediment–water interface (12), a process referred to as electrogenic sulfur oxidation (e-SOx) (15). In laboratory experiments, e-SOx has been shown to exert a strong impact on sedimentary iron and sulfur cycling, leading to a conversion of iron sulfides into iron oxides (16). Here we demonstrate that the same interconversion process of iron minerals occurs in the sediments of a seasonally hypoxic marine basin, and that the large pool of iron oxides can act as a “firewall,” which can substantially delay the development of euxinia.

Significance

Seasonal hypoxia is increasing in coastal areas worldwide, as more nutrients are delivered to the coastal ocean and water temperatures are rising due to climate change. Hypoxia reaches a particularly harmful stage when sulfide, which is highly toxic for marine life, is released to the bottom water. Here, we document a natural microbial mechanism that counteracts the release of free sulfide, thus preventing the most adverse stage of seasonal hypoxia. Electricity-generating cable bacteria produce a large pool of oxidized sedimentary iron minerals, which efficiently bind free sulfide. As cable bacteria are likely abundant in many seasonally hypoxic basins worldwide, their “firewall” mechanism may be widespread.

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Results and Discussion

Response of Pore Water Geochemistry to Seasonal Oxygenation. The sediment geochemistry, sediment fauna, and sedimentary microbial communities were surveyed in Marine Lake Grevelingen (MLG, The Netherlands), a coastal water body (salinity ~ 30) with restricted water exchange with the open North Sea. Over the last decade, MLG has experienced a regular pattern of summer stratification and bottom water oxygen depletion (Fig. S1A), which was also observed in monthly sampling campaigns performed throughout 2012 (Fig. S1B). Bottom water oxygen concentrations (Winkler method; Fig. 1A) were near air saturation in winter and early spring, started to decline in April at the onset of stratification, became hypoxic ($O_2 < 63 \mu\text{M}$) by the end of May, and declined below the detection limit in August ($O_2 < 1 \mu\text{M}$; anoxia). In September, the overturning of the water column resulted in a reoxygenation of the bottom water.

Due to sediment focusing, the deeper basins in MLG experience a strong accumulation ($\sim 2 \text{ cm}\cdot\text{y}^{-1}$) of dark, organic-rich, fine-grained sediment (14). Free sulfide ($\Sigma\text{H}_2\text{S} = [\text{HS}^-] + [\text{H}_2\text{S}]$) accumulates in the pore water to high levels ($\sim 2 \text{ mM}$ at 10 cm depth; Fig. S2), suggesting that intensive organic mineralization takes place in the surface sediment and that sulfate reduction is the major mineralization pathway (estimated to be $\sim 30 \text{ mmol}$

$\text{S}\cdot\text{m}^{-2}\cdot\text{d}^{-1}$ in August; Supporting Information). The burial of pyrite and iron sulfides only scavenges $\sim 39\%$ of the sulfide production (Supporting Information), and bottom water concentrations of nitrate (the alternative electron acceptor) are generally low in MLG (Fig. S1B). Accordingly, oxygen appears to be the main electron acceptor for the oxidation of the large amount of free sulfide that is produced by sulfate reduction (Supporting Information).

Nevertheless, microsensors profiling revealed that O_2 and H_2S were almost never in direct contact in the pore water, as a well-developed suboxic zone was present throughout most of the year (Fig. 1B), i.e., a distinct sediment horizon where neither O_2 nor $\Sigma\text{H}_2\text{S}$ were present in detectable concentrations. This suboxic zone was widest in the first part of the year (annual maximum of $17.6 \pm 4.6 \text{ mm}$ in April), and its width decreased in a stepwise fashion by $\sim 50\%$ in late spring (May, 8.9 ± 2.1 ; June, $7.0 \pm 5.2 \text{ mm}$). When the oxygen saturation of the bottom water dropped below 11% (corresponding to $29 \mu\text{mol}\cdot\text{L}^{-1}$; July and August, Fig. 1A), the H_2S front started to move toward the sediment surface (Fig. 1B). However, free sulfide remained undetectable in bottom water samples collected at 2 m above the sediment surface during the stratified season ($\Sigma\text{H}_2\text{S} < 0.2 \mu\text{M}$; June, August). We additionally determined the $\Sigma\text{H}_2\text{S}$ concentration in the overlying $\sim 10 \text{ cm}$ water of retrieved sediment cores, which confirmed that euxinia did not occur. After the overturning of the basin, the oxygen penetration depth (OPD) was small ($0.7 \pm 0.1 \text{ mm}$), indicative of strong oxygen uptake caused by reoxidation of reduced compounds. September was also the only month when the pore water depth profiles O_2 and H_2S showed an overlap (Fig. 1B), allowing the direct aerobic oxidation of sulfide ($\text{H}_2\text{S} + 2\text{O}_2 \rightarrow \text{SO}_4^{2-} + 2\text{H}^+$). A small suboxic zone reappeared in October ($2.2 \pm 0.9 \text{ mm}$) and gradually expanded ($12.2 \pm 1.6 \text{ mm}$ in December).

Mechanisms of Suboxic Zone Formation. To elucidate the underlying mechanisms of sulfide oxidation and suboxic zone formation, we developed a pH typology for three known mechanisms of aerobic sulfide oxidation (Fig. S3 and Supporting Information): (i) the e-SOx metabolism of the recently discovered cable bacteria (12, 13, 16); (ii) the cycling of iron between reduced and oxidized mineral forms, which is crucially dependent on solid-phase mixing (17, 18); and (iii) the respiratory metabolism of nitrate-accumulating *Beggiatoaceae* (19–22). Each of these three pathways is associated with a characteristic pH depth profile (Fig. 2, Fig. S3, and Supporting Information), which reveals when these mechanisms dominate the pore water geochemistry and are responsible for the formation of the suboxic zone (Fig. 1C). This pH typology predicted that e-SOx by cable bacteria was dominant from January to April, whereas metal cycling dominated in May and June, and respiration of nitrate-accumulating *Beggiatoaceae* created the suboxic zone in fall, after the ventilation and reoxygenation of the bottom water (Fig. 1C).

This temporal succession of sulfur oxidation pathways as predicted by the pH typology analysis was confirmed by direct microscopic observation of microbial and macrofaunal communities. Cable bacteria were enumerated on a seasonal basis by Fluorescence In Situ Hybridization (FISH), which revealed that cable bacteria were abundant in March and May (filament density $402\text{--}480 \text{ m}\cdot\text{cm}^{-2}$, biovolume $1.8\text{--}2.5 \text{ mm}^3\cdot\text{cm}^{-2}$), but remained below the detection limit in August and November (Fig. 1D). When present, cable bacteria were found throughout the upper 40 mm of the sediment, with the maximum density near the sediment surface, high densities in the suboxic zone, and low densities declining into the sulfidic zone (Fig. 2A, March). The filament diameter did not differ significantly (t test; $n = 15$, $P = 0.29$) between March ($1.3 \pm 0.3 \mu\text{m}$) and May ($1.2 \pm 0.3 \mu\text{m}$), suggesting that cable bacteria populations were phenotypically similar.

Quantitative microscopic enumeration of *Beggiatoaceae* was conducted each month (Fig. 1D), and showed low densities in spring (biovolume B: $0.02\text{--}0.05 \text{ mm}^3\cdot\text{cm}^{-2}$), when only a few filaments were found dispersed throughout the upper 2 cm of sediment, and a

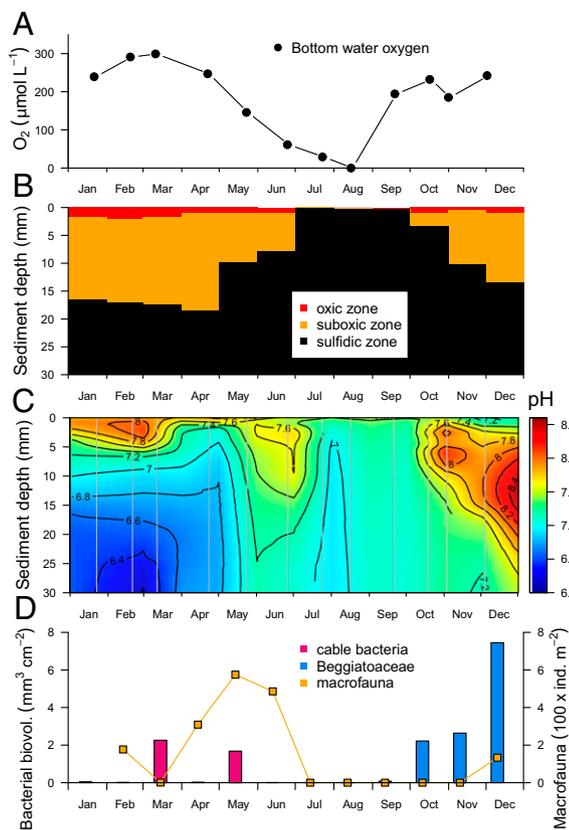


Fig. 1. (A) Oxygen concentration (in micromoles per liter) in the bottom water, measured at 1 m above the sediment surface throughout 2012. (B) Depth distribution of the oxic zone (red), suboxic zone (brown), and sulfidic zone (black) within the sediment throughout 2012. The data were obtained from microsensors measurements in intact sediment cores. (C) Pore water pH values in the upper 30 mm of the sediment measured monthly during 2012, in intact sediment cores. Gray lines indicate sampling dates. (D) Biovolume (in cubic millimeters per square centimeter) of cable bacteria (pink bars) and *Beggiatoaceae* (blue bars), and macrofauna abundance ($100 \times \text{individuals}\cdot\text{m}^{-2}$; orange line) recorded throughout 2012. Cable bacteria were enumerated in March, May, August, and November; *Beggiatoaceae* were counted monthly over the entire year 2012; macrofauna were determined monthly from February to December.

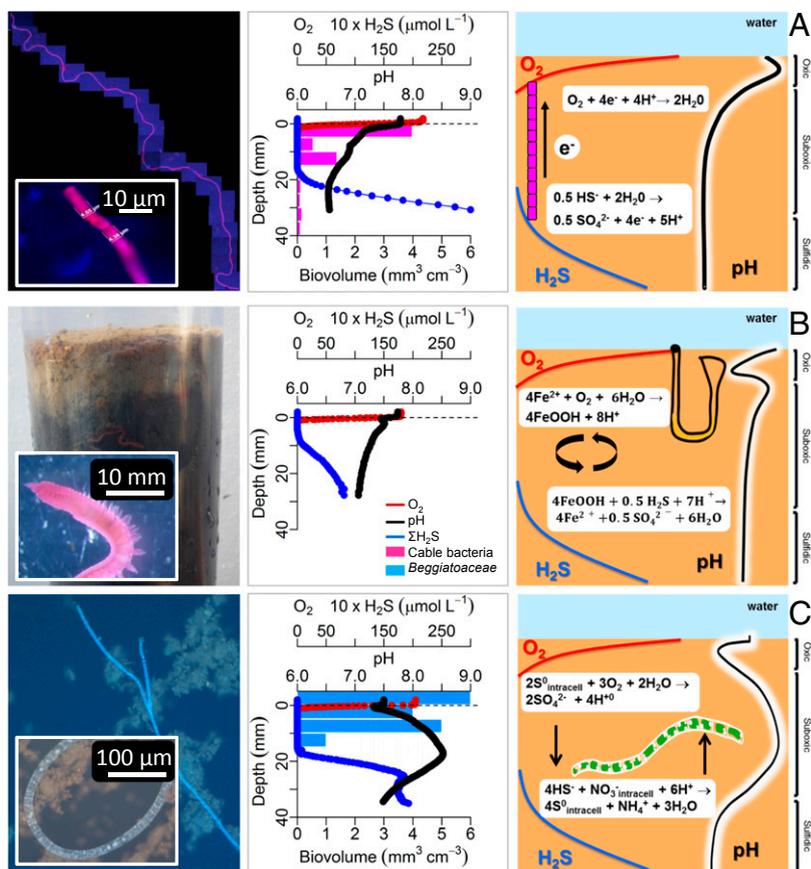


Fig. 2. Three mechanisms of suboxic zone formation were active during 2012 at the study site. (A) FISH image of cable bacterium (DSB706 probe) and close-up of a cable bacterium (Left); depth profiles of O₂ (in micromoles per liter, red line), pH (black line), and ΣH₂S (in micromoles per liter, blue line) and biovolume (in cubic millimeters per square centimeter) of cable bacteria (pink bar) in March (Middle); and schematic representation of geochemical signature and e-SOx carried out by cable bacteria (Right). (B) Sediment core picture and microscope picture of the polychaete *Scoloplos armiger* (Left); microsensor depth profiles in May (Middle); and schematic view of effect of the macrofauna on sedimentary iron and sulfur cycling (Right). (C) *Beggiatoaceae* filament image and close-up of a filament (light microscope, stained with DAPI) (Left); microsensor depth profiles and *Beggiatoaceae* biovolume (in cubic millimeters per square centimeter, pink bar) in December (Middle); and schematic view of geochemical signature and sulfur oxidation carried out by *Beggiatoaceae* (Right). *Scoloplos armiger* image courtesy of Monitoring Task Force (Royal Netherlands Institute for Sea Research).

virtual absence from May to August ($B \leq 0.001 \text{ mm}^3 \cdot \text{cm}^{-2}$). In September, a population of thin *Beggiatoaceae* filaments (diameter d : 2.4 μm; mean filament length L : 70 μm; B : $0.08 \text{ mm}^3 \cdot \text{cm}^{-2}$) was found concentrated right at the O₂–H₂S interface, which likely catalyzed the direct aerobic oxidation of free sulfide. Laboratory studies have estimated that sulfide oxidation at the oxic–anoxic interface by *Beggiatoaceae* may be up to 3 times faster than the autocatalytic aerobic chemical oxidation of sulfide (23–25), thus enabling an efficient competition with the abiotic pathway. From October onward, the biovolume of *Beggiatoaceae* filaments drastically increased (B : $2.2\text{--}7.5 \text{ mm}^3 \cdot \text{cm}^{-2}$), and the depth distribution of the *Beggiatoaceae* closely tracked the progressive widening of the suboxic zone (Fig. 2C). In addition, the filament diameter and length increased significantly compared with September (t test; $n = 672$, $P < 0.001$, regarding both length and diameter), suggesting that a different population of *Beggiatoaceae* was active in late fall that had a metabolism based on intracellular nitrate respiration. Bacterial cell-lysing experiments confirmed that the large *Beggiatoaceae* filaments observed at the field site were storing nitrate into intracellular vacuoles (Fig. S4 and Supporting Information).

The pH typology analysis predicted that suboxic zone in May and June was no longer formed by cable bacteria but that metal cycling caused the separation of O₂ and H₂S horizons in the upper first centimeter of the sediment (Fig. 2B). This coincided with a sharp rise in the abundance and diversity of the macrofauna in the surface sediment (Fig. 1D), suggesting that bioturbation could provide the sediment mixing needed to sustain the metal cycling. An alternative explanation would be that mixing via sediment resuspension is specifically intense during May and June, which is unlikely, however, as meteorological conditions at the field site are typically calm in early summer. With the onset of hypoxia in late June, the macrofauna vanished abruptly, and the sediment remained devoid of macrofauna until December, when recolonization started,

although population densities remained low throughout winter. Upon recolonization in spring, the fauna was dominated by small polychaetes and juvenile bivalves, which only have a shallow burrowing depth, consistent with the limited suboxic zone of 7–9 mm observed in May and June. Because fauna are highly sensitive to free sulfide (9–11), the deep removal of sulfide by cable bacteria in early spring may have promoted faunal recolonization.

Microbial Competition for Reduced Sulfur Compounds. In MLG, we observed that cable bacteria were dominant throughout spring, whereas sulfur oxidation was largely carried out by nitrate-accumulating *Beggiatoaceae* throughout fall. This pattern was not only observed in 2012, when detailed monthly sampling was conducted, but was confirmed by seasonal surveys over the period 2011–2015 (Fig. 3 and Supporting Information). Combining microsensor profiling and pH-signature analysis, we found that the geochemical signature of cable bacteria is significantly more present in spring, whereas the activity of *Beggiatoaceae* is more likely encountered in fall (Fig. 3). For example, in spring 2015, all sampled sediment sites below 15 m water depth showed the cable bacteria signature, whereas, in fall 2011 and 2014, all sediment revealed the geochemical signature of nitrate accumulating *Beggiatoaceae*. This implies that two distinct types of filamentous S-oxidizing bacteria were competing for the same geochemical niche, but that each type was competitively successful during a distinct period of the year.

The development of *Beggiatoaceae* after summer suggests a better survival of the anoxic period, which could be due to the use of nitrate as an alternative electron acceptor to oxygen. Although both cable bacteria (26) and *Beggiatoaceae* (19–22) can use nitrate for respiration, cable bacteria reach lower population densities when nitrate is the sole electron acceptor (26). Moreover, the nitrate concentration in the bottom water was low ($< 1.7 \mu\text{M}$) during summer (Fig. S1B), and thus, nitrate reduction was likely

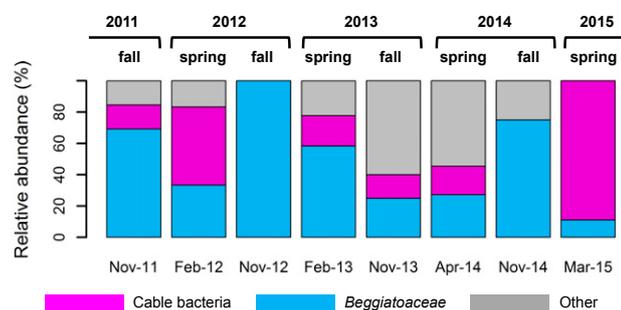


Fig. 3. Relative abundance of the dominant sulfur oxidation pathways carried out by cable bacteria (pink bar), *Beggiatoaceae* (blue bar), and other processes (gray bar) during each fall and spring from November 2011 to March 2015. The geochemical signature (characteristic depth profiles of O_2 , pH, and H_2S recorded by microsensors profiling) was used to determine the dominant pathway of sulfur oxidation at any given time at different sites within MLG (*Supporting Information*).

insignificant in sustaining microbial metabolism during anoxia. However, nitrate accumulation before anoxia could have played a role. Presently, there are no indications that cable bacteria can accumulate electron acceptors, whereas *Beggiatoaceae* can store nitrate in intracellular vacuoles (19–22, 25), which can be used as an electron acceptor reservoir to survive the summer period of low bottom water oxygenation (25).

Our seasonal surveys indicate that cable bacteria replace the *Beggiatoaceae* population in winter, yet the reasons for this population switch are not fully understood. One intriguing question is how cable bacteria can “invade” a sediment where a suboxic zone is already established by *Beggiatoaceae*, as recent laboratory experiments show that cable bacteria filaments progressively extend downward from an initial overlap of oxygen and sulfide near the sediment–water interface (27–29). A preexisting suboxic zone thus poses a barrier for sediment colonization by cable bacteria. Future research should hence clarify the drivers and controls of what appears to be a yearly recurrent, and hence predictable, switch between two groups of filamentous S-oxidizing bacteria.

Impact of Cable Bacteria on Geochemistry. Cable bacteria and nitrate-accumulating *Beggiatoaceae* are both capable of efficient sulfide oxidation leading to the creation of a wide suboxic zone. Hence, with regard to biogeochemical cycling in seasonally hypoxic basins, one could ask to what extent it matters whether one or the other is the dominant sulfur oxidizing microbial population? Solid-phase data collected at the field site reveal a notable difference in the iron mineral phases between spring and fall (Fig. 4A), and suggest that cable bacteria induce a strong seasonal iron cycling. In March, when cable bacteria were active, a strong depletion of Acid Volatile Sulfides (AVS, interpreted to be mostly iron monosulfides, FeS) occurred in the suboxic zone ($11.9 \pm 7.3 \mu\text{mol}\cdot\text{S}\cdot\text{g}^{-1}$ over the first 3.3 cm; Fig. 4A), compared with high values in November ($117.0 \pm 26.4 \mu\text{mol}\cdot\text{S}\cdot\text{g}^{-1}$ over the first 3.3 cm; Fig. 4A), when nitrate-accumulating *Beggiatoaceae* were abundant. Exactly the opposite trend was seen in the extractable iron (hydr)oxides (FeOOH), which showed a much higher accumulation in spring ($163 \pm 52 \mu\text{mol}\cdot\text{Fe}\cdot\text{g}^{-1}$; Fig. 4A) than in fall ($99 \pm 27 \mu\text{mol}\cdot\text{Fe}\cdot\text{g}^{-1}$). Together with our microsensors and microscopy data, these solid-phase data suggest the following seasonal iron cycle (Fig. 4B): (i) conversion of FeS to FeOOH in spring by cable bacteria; (ii) the downward mixing of the FeOOH-rich surface sediment layer in late spring by the newly colonizing fauna, thus enhancing the observed S oxidation by metal oxide reduction; (iii) the conversion of FeOOH back to FeS when free sulfide rises to the sediment–water interface during the summer hypoxia period; and (iv) the persistence of an FeS pool in the suboxic zone during fall when

nitrate-accumulating bacteria are active. Alternative mechanisms, such as winter resuspension events or a strong sedimentation of allochthonous iron (hydr)oxides in winter, cannot suitably explain the observed conversion of FeS to FeOOH in spring (*Supporting Information*). Accordingly, we conclude that observed seasonal iron cycling is principally driven by the FeOOH-forming activity of cable bacteria in spring, and this iron cycle would not occur if nitrate-accumulating *Beggiatoaceae* were dominant throughout the year. Our study therefore demonstrates that not only external environmental factors, such as bottom water oxygen availability, are driving the sedimentary iron and sulfur cycling in seasonally hypoxic basins but that the intrinsic population dynamics of the microbial community, and particularly rapid shifts in sulfur oxidizers, can be equally important.

The metabolic activity of cable bacteria exerts a profound impact on the sediment geochemistry through its effect on pore water pH (15). The electrogenic metabolism induces a spatial uncoupling of sulfide oxidation and oxygen reduction (Fig. 2A), and, accordingly, the production and consumption of protons occur widely segregated in space. The establishment of acidic conditions within the deeper suboxic zone promotes the dissolution of iron sulfides, which provides an extra H_2S supply to the cable bacteria in addition to sulfate reduction (16), and the oxidation of this extra H_2S generates more protons, thus establishing a positive feedback (15). In laboratory experiments, it has been shown that FeS dissolution provides up to 40–94% of the sulfide for e-SOx (15, 16), and can completely exhaust the sedimentary FeS pool over a period of weeks (27), while generating a surface enrichment of FeOOH. The observed FeS depletion in the top 4 cm in March shows that cable bacteria also induce strong FeS dissolution in the field, and we speculate that the depletion of the sedimentary FeS stock (Fig. 4A) may have limited the electron donor supply, causing the demise of the cable bacteria population in late spring (30).

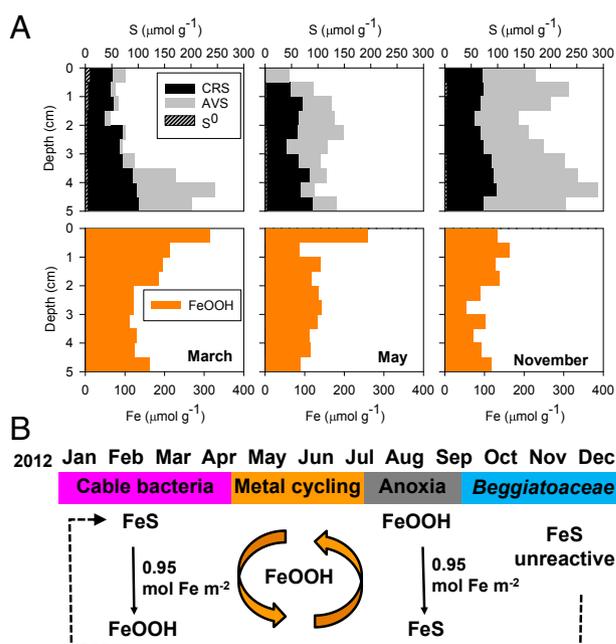


Fig. 4. (A) Solid-phase geochemistry in the top 5 cm of the sediment. Depth profiles of S^0 , CRS, and AVS (in micromoles S per gram, *Top*) and iron (hydr)oxides (in micromoles Fe per gram, *Bottom*) measured in March (*Left*), May (*Middle*), and November (*Right*). The iron (hydr)oxides represent the total of the nonsulfidized iron. (B) Illustration of interconversion between FeS and FeOOH during four time periods in 2012, when the sediment geochemistry was dominated by cable bacteria, metal cycling promoted by sediment mixing, anoxia, and *Beggiatoaceae*.

Cable Bacteria in Seasonally Hypoxic Systems and Euxinia. Conventionally, the formation of sulfidic bottom waters is considered to be closely linked to the exhaustion of energetically favorable electron acceptors in the water column, such as oxygen and nitrate (1). Once these electron acceptors are depleted in the bottom water, the oxidation of sulfide in the surface sediment layer is halted, thus enabling a release of sulfide to the overlying water. However, laboratory sediment incubations (31, 32) have previously shown that the disappearance of oxygen and nitrate (anoxia) is not necessarily synchronous with the appearance of free sulfide (euxinia). In sediments that contain a large pool of reactive iron (hydr)oxides before the onset of anoxia, this pool can act as a firewall against the release of free sulfide from the sediment, thus delaying the onset of euxinia (31, 32). Iron (hydr)oxides have a high binding capacity for free sulfide (33); therefore, the efflux of free sulfide only starts after exhaustion of the iron (hydr)oxide pool, which can delay euxinia by several weeks, depending on the size of the initial reactive iron pool (31, 32).

At present, this iron oxide-mediated firewall mechanism has not been demonstrated to occur in seasonally hypoxic basins. Moreover, it is also not expected to occur, as the mechanism requires a yearly buildup of an iron (hydr)oxides pool before the onset of summer anoxia, which is not obvious in seasonally hypoxic environments. Coastal sediments typically accumulate sizeable pools of iron (hydr)oxides only when subjected to strong levels of bioturbation by infauna (34). The ventilation of macrofaunal burrows with oxygen-rich overlying water promotes the oxidation of dissolved ferrous iron (Fe^{2+}) in the pore water (35), whereas particle reworking enhances the oxidation of deeply buried iron sulfides by transporting them upward to the oxic zone of the sediment (36). However, as shown here, such large and deep-burrowing fauna are typically absent from seasonally hypoxic sediments, because the yearly recurrent oxygen depletion increases mortality and decreases recruitment success (11, 37). For this reason, the sediments of seasonally hypoxic coastal systems are generally thought to have a low buffer capacity toward the release of free sulfide (38).

Although sampling at monthly resolution is not sufficient to accurately document the magnitude of the time lag, we observed that anoxia did not coincide with euxinia. Even when the bottom water was devoid of oxygen in August, and nitrate was fully depleted, no free sulfide (concentrations were below the detection limit of $< 0.2 \mu\text{M}$) was detected in the bottom water. Two potential mechanisms could explain this delay in the formation of euxinia relative to anoxia. The presence of a deep suboxic zone before the onset of bottom water anoxia is one potential mechanism. Before free sulfide can escape the sediment, the suboxic zone has to be transiently replenished with free sulfide, either through local production of sulfide via sulfate reduction or via diffusion of sulfide from deeper sediment horizons. As both cable bacteria and *Beggiatoaceae* induce a suboxic zone, they both induce this form of euxinia delay. Analysis of the curvature of the $\Sigma\text{H}_2\text{S}$ depth profile provides an estimate of the sulfide production rate of $0.3 \text{ mol}\cdot\text{H}_2\text{S}\cdot\text{m}^{-3}\cdot\text{d}^{-1}$ within the first 10 cm of sediment. Accordingly, ~ 6 d are required for sulfide to accumulate in the suboxic zone up to the $\Sigma\text{H}_2\text{S}$ ~ 2 -mM level observed in August (porosity of 0.9). In reality, this accumulation of sulfide in the suboxic zone will proceed faster, as molecular diffusion also supplies free sulfide from beneath the suboxic zone. Accordingly, the formation of a deep suboxic zone only delays euxinia for a short period, on the order of a few days at most.

The iron oxide-mediated firewall mechanism is a more effective mechanism for H_2S removal, and has been previously shown to delay sulfide effluxes from sediments for a period of weeks (31, 32). Our results show that cable bacteria generated a large pool of reactive iron (hydr)oxides before the onset of summer hypoxia ($0.95 \text{ mol}\cdot\text{Fe}\cdot\text{m}^{-2}$, as calculated from the difference in FeS and FeOOH inventories over 0–4 cm between spring and fall; *Supporting Information*). Given a depth-integrated sulfide

production rate of $30 \text{ mmol}\cdot\text{H}_2\text{S}\cdot\text{m}^{-2}\cdot\text{d}^{-1}$, this iron (hydr)oxide pool could potentially buffer H_2S production up to ~ 36 d. This period is considerably longer than the observed period of anoxia at the study site in 2012 (< 20 d), and hence may explain the observed absence of bottom water euxinia in August 2012.

Conclusion

Cable bacteria have only recently been discovered (12, 13), and, hence, little is known about their ecology, life cycle, and natural distribution. Our results demonstrate that cable bacteria can have a major impact on sedimentary biogeochemical cycling in a seasonally hypoxic basin, with potential basin-scale impacts on water column chemistry. In a first report on the occurrence of cable bacteria under natural conditions, it was demonstrated that cable bacteria thrive globally in a wide range of marine sediment habitats, such as coastal mud plains and salt marshes, but that they particularly seem abundant in seasonally hypoxic basins (14). If cable bacteria in other coastal systems follow a similar seasonal cycle to that of MLG, the iron oxide firewall mechanism proposed here could be widely prevalent, and may explain the relatively rare reports of euxinia in coastal systems affected by seasonal hypoxia. However, to accurately document the magnitude and efficiency of this buffer mechanism, a comparison of the sediment geochemistry and microbiology is needed across multiple seasonally hypoxic systems at a higher-than-monthly resolution. Such investigations are crucial, given that seasonal hypoxia in coastal areas is increasing worldwide due to anthropogenic nutrient input and climate change (1).

Materials and Methods

Sampling. We performed monthly sampling campaigns on the *R/V Luctor* in 2012, in the seasonally hypoxic MLG (39). Investigations took place in the Den Osse basin, a deep gully located in the southwestern part of the lake (maximum water depth 34 m; 51.747°N , 3.890°E), and we examined the water column chemistry and sediment biogeochemical processes. Discrete bottom water samples were collected with a 12-L Niskin bottle to assess the O_2 and H_2S concentrations. Water samples from the Niskin bottle were collected via gas-tight Tygon tubing. Bottom water oxygen concentrations were measured using an automated Winkler titration procedure with potentiometric endpoint detection (Mettler Toledo DL50 titrator and a platinum redox electrode). Bottom water $\Sigma\text{H}_2\text{S}$ concentrations were determined spectrophotometrically (40). Intact sediment cores (6 cm \varnothing) were retrieved with a UWITEC gravity corer in triplicates. All cores were inspected on retrieval, and only undisturbed cores were used for measurements. Immediately after collection, sediment cores were transported to a nearby laboratory, where microprofiling was started within 2 h of collection and conducted under climate-controlled conditions (temperature of in situ bottom water).

Microsensor Profiling. Microsensor profiling was performed using commercial microelectrodes (Unisense A.S.) for O_2 (25- or 50- μm tip; Unisense), pH (200- μm tip diameter), and H_2S (50- μm tip diameter). Oxygen microprofiles were made at 25- to 50- μm resolution, with a two-point calibration made in air-saturated seawater (100% saturation) and at depth in anoxic sediment (0% saturation). For H_2S and pH, depth profiles were made at 200- μm resolution in the oxic zone, and 400- or 600- μm resolution below. Calibrations for pH were made with three National Bureau of Standards (NBS) standards and a Tris buffer to correct for salinity effects, and pH is reported on the total scale. For H_2S , a five-point calibration was made using Na_2S standards. Total free sulfide ($\Sigma\text{H}_2\text{S} = \text{H}_2\text{S} + \text{HS}^-$) was calculated from H_2S based on pH measured at the same depth using the R package AquaEnv (41). The OPD is operationally defined as the depth below which $[\text{O}_2] < 1 \mu\text{M}$, and the sulfide appearance depth (SAD) is operationally defined as the depth below which $[\text{H}_2\text{S}] > 1 \mu\text{M}$. The diffusive oxygen uptake (DOU) was calculated from the oxygen depth profiles as in ref. 14.

Pore Water and Solid-Phase Geochemistry. Pore water was extracted from the sediment using centrifugation (15 min at $4,500 \times g$) and was filtered (0.45 μm) and subsampled under N_2 . Centrifuged pore water was subsampled for sulfide, where samples were fixed with zinc acetate and stored at 4°C . Sulfide was measured spectrophotometrically (40). Centrifuged sediment samples were freeze-dried, then ground in an N_2 -purged glove box. Sediment sulfur fractions were separated using the extraction described in ref. 42, and modified as in ref. 43 to include an extraction step for elemental

sulfur. AVS and chromium reducible sulfide (CRS) were quantified using iodometric titration. Solid-phase Fe phases were also extracted and separated according to ref. 44, where Fe oxides were measured as the total of the nonsulfidized Fe pools.

Microscopy. *Beggiatoaceae* filaments were identified via inverted light microscope (Olympus IM) within 24 h of retrieval. Intact sediment cores were sectioned at 5-mm intervals over the top 4 cm from which subsamples (20–30 mg) were used to count living *Beggiatoaceae* filaments. The biovolume was determined by measuring length (10 \times) and width (40 \times) of all filaments found in the subsample, according to ref. 45. Microscopic identification of cable bacteria was achieved by FISH, using a *Desulfobulbaceae*-specific oligonucleotide probe (DSB706; 5'-ACC CGT ATT CCT CCC GAT-3'), according to Schauer et al. (27). The depth distribution of cable bacteria was quantified in March, May, August, and November 2012. Cable bacteria biovolume per unit of sediment volume (in cubic millimeters per cubic centimeter) was calculated based on measured filament length and diameter, as well as the areal biovolume of cable bacteria (in cubic millimeters per square centimeter) by

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